

A Biogenerated Polymetallic Catalyst from Society's Wastes

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Authors' contributions

This work was carried out in collaboration among all authors. All authors read and approved the final manuscript.

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ABSTRACT

Aims: Preparation of the new metals-polymeric composite, Met_x-EPS (I), to be used as a green catalyst in water or in two-phase aqueous conditions.

Study Design: Recovery and valorization of polymetallic wastes to obtain directly new catalysts using a microorganism to explore their application in removal of difficult and dangerous chemical pollutants present in aqueous environment.

Place and Duration of Study: Dipartimento di Scienze Molecolari e Nanosistemi, Università Ca' Foscari Venezia, Venezia Mestre, Italy; University of Nova Gorica, Nova Gorica, Slovenia, Institut Jozef Stefan, Ljubljana, Slovenia and Department of Biology, Biotechnical Faculty, University of Ljubljana, Ljubljana, Slovenia, between February 2018 and January 2019.

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Methodology: For the preparation of $\text{Met}_x\text{-EPS}$ (I), the metals source was an exhausted catalytic converter that was grinded and treated with an acidic solution at 20-25°C. After filtration, the solution was concentrated, neutralized and added to a broth of *Klebsiella oxytoca* DSM 29614 to produce (I) where metals are embedded in a peculiar polysaccharide structure. The composite was easily recovered from the fermentation broth and purified. The process protocol was verified many times and was shown to be reproducible satisfactorily. The % recovery of metals, originally present in the converter, was good as determined by atomic absorption. The morphology and the chemical state of main metals in (I) were investigated by X-ray absorption spectroscopy methods (XANES and EXAFS). No metallic alloy seems to be evident.

Results: As first application of (I) as catalyst, the hydrodechlorination treatment of polychlorinated biphenyls (PCBs) was studied in water/methanol. A significant removal of higher chlorinated congeners was observed working at 1MPa of hydrogen and 60°C. This result improves significantly and surprisingly the methodology, previously studied by us using mono- or bi-metals embedded in the same polysaccharide moiety, indicating that positive synergies among the different metals were operating.

Conclusion: The preparation of this new polymetallic species embedded in a polysaccharide moiety starting from spent catalytic converters represents an alternative valorisation of metallic wastes. Moreover, a synergic effect was exerted by the different metals when the catalyst $\text{Met}_x\text{-EPS}$ (I) was used in the hydrodechlorination treatment of polychlorinated biphenyls (PCBs) in water/methanol. Finally, a promising preliminary proof of concept for the removal of polychlorinated aromatic pollutants even in contaminated aqueous sites was carried out.

Keywords: *Metals-polymeric composite; biogenerated polymetallic exopolysaccharide; new catalyst from metallic wastes; hydrodechlorination of PCBs in water.*

1. INTRODUCTION

As now requested by the circular economy, valuable metals have to be recovered from industrial scraps, such as for example exhausted catalytic converters. Current approaches to treat relevant metallic waste categories are based on pyrometallurgical, hydrometallurgical or bio-hydrometallurgical methods, which present advantages and disadvantages in terms of energy consumption, use of special equipments and difficulty in the purification [1,2]. Then a following treatment is necessary to produce metallic catalysts. To overcome these drawbacks and to obtain directly a composite useful as catalyst, our protocol at first treats grinded exhausted catalytic converters at room temperature with a concentrated acidic solution, such as aqua regia or 80% nitric acid, to dissolve metals as ions and, after filtration of the solid residue, concentration and neutralization of the solution, takes advantage of the property of a plurimetal resistant microorganism *Klebsiella oxytoca* DSM 29614. This microorganism is able to generate, in the presence of heavy metal cations, a specific capsular exopolysaccharide (EPS) which may embed them and then be extruded from the cell so to be easily purifiable [3-4]. So a new metals-polymeric composite, $\text{Met}_x\text{-EPS}$ (I) was obtained from exhausted catalytic converters and, after characterization

with different methods, its catalytic activity and a possible synergic effect due to the presence of different metals embedded in EPS was verified in the hydrodechlorination treatment of a methanolic solution of polychlorinated biphenyls (PCBs) in distilled water and in a sample of polluted sea water (Mar Piccolo, Taranto, Italy). The results were compared with those recently obtained by us using Pd-EPS or Pd,Fe-EPS [5].

2. MATERIALS AND METHODS

2.1 Materials and Instrumentation

Nutrient broth (Difco), n-hexane (pesticide grade; Romil), PCB standards (Accu Standard) were utilised. Aroclor 1260 and the other reagents were Sigma-Aldrich products and used as received. The total amounts of elements were determined versus their relative standards solutions by inductively coupled plasma atomic emission spectrometry (ICP-AES) (Optima 3100, Perkin Elmer). The FT-IR spectra (KBr pellets) of $\text{Met}_x\text{-EPS}$ were recorded on an FT-IR Nicolet Magna 750 instrument. X-ray absorption spectra were measured on of $\text{Met}_x\text{-EPS}$ sample in the energy range of K or L3 absorption edges of constituent metal cations (Pd K-edge (24350 eV), Rh K-edge (23220 eV), Zr K-edge (17998 eV), Pt L3 edge (11564 eV), Ce L3 edge (5724 eV)) to obtain corresponding XANES (X-ray Absorption

Near Edge Structure) and EXAFS (Extended X-ray Absorption Fine Structure) spectra of the investigated elements. The $\text{Met}_x\text{-EPS}$ sample was prepared in the form of homogeneous self-standing pellets, each optimised to provide the total absorption thickness (μd) of about 1.5 above the investigated absorption edges. All absorption spectra were measured at room temperature in transmission detection mode at two different synchrotron radiation facilities (Zr, Pt and Ce at XAFS beamline at Elettra in Trieste, Italy; Pd and Rh at BL22 (Claess) beamline of ALBA in Barcelona, Spain); a Si (111) double crystal monochromator was used at the XAFS beamline of the Elettra and a Si(311) at BL22 of ALBA. The intensity of the monochromatic X-ray beam was measured by three consecutive ionization detectors. The samples were placed between the first pair of detectors. The exact energy calibration was established with simultaneous absorption measurement on the corresponding reference metal foil placed between the last pair of detectors. Absolute energy reproducibility of the measured spectra was ± 0.03 eV. The absorption spectra were measured within the interval [-150 eV to 1000 eV] relative to the investigated absorption edge. In the XANES region, equidistant energy steps of 0.5 eV were used, while for the EXAFS region, equidistant k-steps ($\Delta k = 0.03 \text{ \AA}^{-1}$) were adopted, with an integration time of 1 s per step. The quantitative analysis of XANES and EXAFS spectra is performed with the IFEFFIT program package [6] in combination with FEFF6 program code [7] for ab initio calculation of photoelectron scattering paths. The determination of PCB congeners was performed in full scan mode on a fused silica capillary column (HP5-MS 30 m, 0.25 mm x 0.25 mm; Agilent Technologies) installed in a ThermoFinnigan (Trace GC 2000) coupled to a quadrupole mass spectrometer (ThermoFinnigan Trace MS), using the same conditions described in our previous work [5].

2.2 Recovery of Acidic Metallic Solutions from Grinded Exhausted Catalytic Converters and Preparation of a New Metals-polymeric Composite (I)

As typical procedure, grinding and pulverizing 20 g of exhausted catalytic converter was carried out using a mortar. The obtained powder was transferred to a 1 liter flask, into which 400 mL of 80% HNO_3 or of aqua regia [65% HNO_3 : 37% HCl (1:3)] was poured under stirring; after 24 h at 20-25°C the mixture was filtered and the residue

solid washed with distilled water; the filtrate was then concentrated at reduced pressure up a volume of 80 mL. This solution was split in two parts, one was used for quantitative analyses of the extracted metals (Table 1), the other (40 mL) was neutralized to pH 7.0 with a concentrated NaOH solution until a final volume of 50 mL and immediately added to a culture of *Klebsiella oxytoca*, DSM 29614 in 1 liter of NaC medium, which contained minimum minerals and sodium citrate as sole carbon and energy source, and had been prepared adapting the protocol already described in our previous works [3,4]. The culture was grown under anaerobic conditions for 5 days and when a cell density upper than 1.0 Abs at 600 nm was reached, 50 ml of instantly buffered plurimetalllic sample were added. After 48 h from this addition, the cell culture was put in a centrifuge to eliminate bacterial cells. The supernatant was treated with 800 ml of cooled 95% ethanol to precipitate $\text{Met}_x\text{-EPS}$. The purification procedure was repeated twice. After drying under vacuum, the material was finally grinded to powder to obtain 1-1.3 g of polymetallic polysaccharide. The preparation of (I) using metals recovered by aqua regia treatment was repeated for three times verifying good reproducibility of the results.

2.3 Element Determination in a New Metals-polymeric Composite (I)

Samples (1 mg) of dry pulverized polymetallic composites were digested with 2 ml of aqua regia, heating the mixture at 60°C until a complete dissolution for element determination. The recovery yield of the main metals is reported in Table 1.

2.4 Optimized Protocol of Hydrodechlorination of Aroclor 1260 Using the New Catalyst (I)

A suitable amount of $\text{Met}_x\text{-EPS}$ (I) to obtain a substrate/Pd $\approx 8/1$ mol ratio (8.8 mg) was stirred in a Schlenk tube under nitrogen in 4 ml of distilled water or 4 ml of Mar Piccolo (Taranto, Italy) water for about 10 min. A solution of 7 mg of Aroclor 1260 (~ 0.0195 mmol assuming MW 358 as medium value) in 2 ml of methanol was then added to the aqueous phase as well as 1.7 mg (0.021 mmol) of ammonium acetate. The Schlenk tube was then transferred into a 150 ml stainless steel autoclave under nitrogen, pressurized with 1 MPa of H_2 , instead of 3MPa used previously with Pd-EPS or Pd, Fe-EPS [5],

and stirred at 60°C for 20 h. The reactor was then cooled to room temperature and opened under nitrogen. Then, the organic products were separated from the catalyst extracting them from the aqueous phase with 10 mL of n-hexane. The procedure of extraction was repeated three times. The organic phases were collected and dried on Na₂SO₄, then concentrated using a nitrogen flow (Turbovap 2, Caliper Science) and analyzed by GC-MS.

3. RESULTS AND DISCUSSION

3.1 Catalyst Characterization

The results of elemental determination are reported in Table 1.

Even if the aqua regia treatment is, as expected, more efficient to extract metals from exhausted catalytic converters, the % of recovery of metals, bonded or encapsulated in the new composite, after the biological treatment, is not significantly affected and is very high (80->99%) with the exception of tungsten (Table 1).

As already described in more detail [8], Met_x-EPS (I) was characterized by FT-IR and by the x-ray absorption spectroscopy methods XANES and EXAFS. The IR spectrum (Appendix, Fig. 1) confirmed the presence of metals bound to capsular EPS. It looks quite similar to the spectrum of a previous prepared Pd-EPS [4] but with a very intense sharp band at 1384 cm⁻¹. The intense symmetric stretching of COO⁻ vibration to 1384 cm⁻¹, on the basis of literature, should demonstrate, in our opinion, that carboxyl group was a contributor in the sorption of metals [9, 10]. In the region between 1550-1650 cm⁻¹ it is possible to observe two peaks, one of them or both could be due to the asymmetric stretching of

COO⁻ vibration bounded to metals [9]. Indeed the EPS is formed by an eptameric structure with two D-glucuronic acids, four L-rhamnose, and one D-galactose [11]. However it is not possible to exclude also contribution of nitrogen species such as any nitrogen heterocyclic compounds eventually formed during the fermentation or membrane protein amide bonds [9]. Finally, the other intense peak at 1065 cm⁻¹, as it was already observed for other similar metal EPS, is to attribute to phosphate groups [4,10].

The presence of metals bounded to capsular EPS and valence and structural information of metal cations in the Met_x-EPS complex at the atomic scale were confirmed by x-ray absorption spectroscopy methods XANES and EXAFS [8]. It is known that, in the XANES spectrum, an increase of oxidation state results in a shift of each absorption feature to higher energies [4,12]. The Pd K-edge profile of the Pd in the Met_x-EPS complex are identical to the Pd XANES spectrum of the Pd metal foil with fcc crystalline structure and also to FePd-EPS(A), FePd-EPS(B) and Pd-EPS samples, previously reported by us [4]. It clearly indicates that palladium also in (I) is predominantly as Pd (0) (Appendix, Fig. 2a).

Results of Pd K-edge EXAFS analysis (Appendix, Fig. 2b) suggested that Pd in (I) is in the form of Pd(0) metallic nano-particles with fcc crystalline structure, similar to that found in mono- and bi-metallic (Pd-EPS and FePd-EPS) samples from previous investigations [4]. However smaller nano-clusters of Pd metal are formed and the average size of Pd metal nanoparticles is less than 1 nm (Appendix, Table 1A). Furthermore, in EXAFS analysis of Pd neighborhood, we found that part of Pd atoms is coordinated to oxygen atoms at about 1.97 Å which is characteristic for Pd oxides (Appendix,

Table 1. Main metals extracted from exhausted catalytic converters and present in Met_x-EPS^a

Analyzed metals	C (mg/L) HNO ₃ (n.a.) sol. (inc. ±10%)	C (mg/L) aqua regia (a.r.) sol. (inc. ±10%)	(Met _x EPS) _{n.a.} metal recovery yield% ^b	(Met _x EPS) _{a.r.} metal recovery yield% ^b
Pd	428	748	89	99
Pt	31.9	479	81	83
Rh	2.07	111	84	85
Al	1623	2413	80	>99
Ce	nd	2158	nd	99
Zr	nd	nd	nd	nd
W	121	174	29	10

^a Experimental conditions: 20 g of exhausted catalytic converter; 400 ml of 80% HNO₃ or aqua regia; T = 20°C; t = 24 h; final volume of the acidic solution after distillation = 80 ml. ^b (Amount of metal encapsulated in EPS)/(amount of metal present in the starting acidic solution)%. nd = not determined

Table 1A). So it is possible that about 10% of Pd atoms is in the form of PdOx nanoparticles simply encapsulated in the polysaccharide moiety or that Pd atoms located on the surface of the Pd metallic nanoparticles may be bonded to the OH or COOH groups of EPS. The results (Appendix, Fig. 2b) are similar, but not identical, to those observed on mono- and bi-metallic samples [4].

The Pt L3-edge XANES analysis would suggest that there are two forms of Pt in (I), one is metallic Pt(0) nanoparticles (about 36 %), as expected, the other in the form of a Pt(II) complex bound to two Cl atoms and two nitrogen atoms like in PtCl₂(pyridine)₂ (Appendix, Fig 3a) [13]. Pt L3-edge EXAFS results strongly support this finding (Appendix, Fig 3b, Table 2A). The results clearly suggest that microorganism *Klebsiella oxytoca* encapsulates Pt species in EPS partly in the form of metallic Pt(0) nanoparticles and partly in the form of complexes structurally similar to PtCl₂(pyridine)₂. It is difficult to have a clear-cut explanation for the latter form. Maybe PtCl₂, originally added to the fermentation broth, can bind to nitrogen heterocyclic species, e.g. piperazine derivatives, probably formed during the fermentation process [14].

The Ce L3-edge XANES analysis (Appendix, Fig. 4) and the comparison with CeVO₄ and CeO₂ XANES spectra showed that all Ce in the Met_x-EPS sample is in the form of three valent Ce cations, most probably in the form of Ce³⁺ oxide nanoparticles [15]. Finally the results of Zr and Rh EXAFS analysis of Met_x-EPS sample (Appendix, Figs. 5 and 6) suggested that both elements, Zr and Rh, are in oxidized form, as a mixture of different Zr and Rh nano-oxides. Zr-Zr distances are characteristic for Zr oxides [16] while Rh-Rh distances are characteristic for crystalline Rh₂O₃ [17]. Also Ce, Zr or Rh cations coordination to OH and COOH groups cannot be excluded.

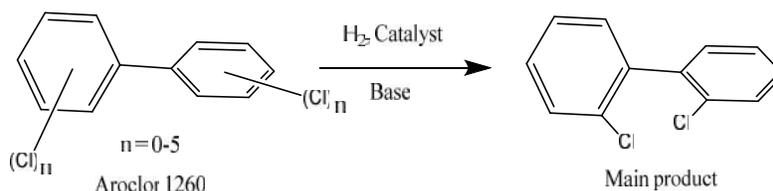
3.2 Catalyst Activity

Removal of polychlorinated pollutants present in water by catalysed hydrodechlorination is still a

challenging target due to the presence of different catalyst poisons. As part of our research activity in environmental remediation and in finding potential sustainable protocols using heterogeneous water compatible catalysts we had recently studied the hydrodechlorination of a methanolic solution of Aroclor 1260 (a reference PCB mixture) with Pd-EPS and Fe₂Pd-EPS in pure distilled water in the presence of ammonium acetate [5] (Scheme 1).

An improvement in the dehalogenation reaction using the bimetallic catalyst had been observed working at a pressure of 3MPa of hydrogen with a substrate/catalyst 8/1 molar ratio (referred to Pd amount), at 60°C in 20 h, but a relevant amount of Cl₅-Cl₇ congeners was still present. Therefore we were encouraged to test this new polymetallic (Met_x-EPS) (I) on the same substrate to try to improve our previous results, working either in pure distilled water or diluting this PCBs mixture in a sample of polluted sea water taken from Mar Piccolo (Taranto, Italy) which contained also other pollutants such as PAHs and heavy metals. Quite surprising, very stimulating and promising results were achieved working under much milder reaction conditions. As shown in Fig. 1a it was possible to obtain a complete removal of congeners having a content of chlorine atoms >3 working at a pressure of 1MPa of hydrogen with the same substrate/catalyst 8/1 molar ratio (referred to Pd amount). Quite satisfactory results were also obtained using a real polluted water spiked with Aroclor 1260 as depicted in Fig. 1b.

It indicates that this new catalyst is quite robust and is able of working in water and of resisting to the poisoning. Synergic positive effects due to the presence of polymetals were clearly demonstrated either in the hydrodechlorination reaction or in the ability to break down/reduce the amounts of palladium poisons present in the polluted sea water. Furthermore PCBs concentrations of some very toxic dioxin-like PCBs congeners: BZ 118, 156, 157, 167, 170, 180, 189 in Aroclor 1260 were reduced to nearly zero (A%) (Table 2) using this new catalyst and it is a significant improvement in the used protocol.



Scheme 1. Hydrodechlorination of Aroclor 1260

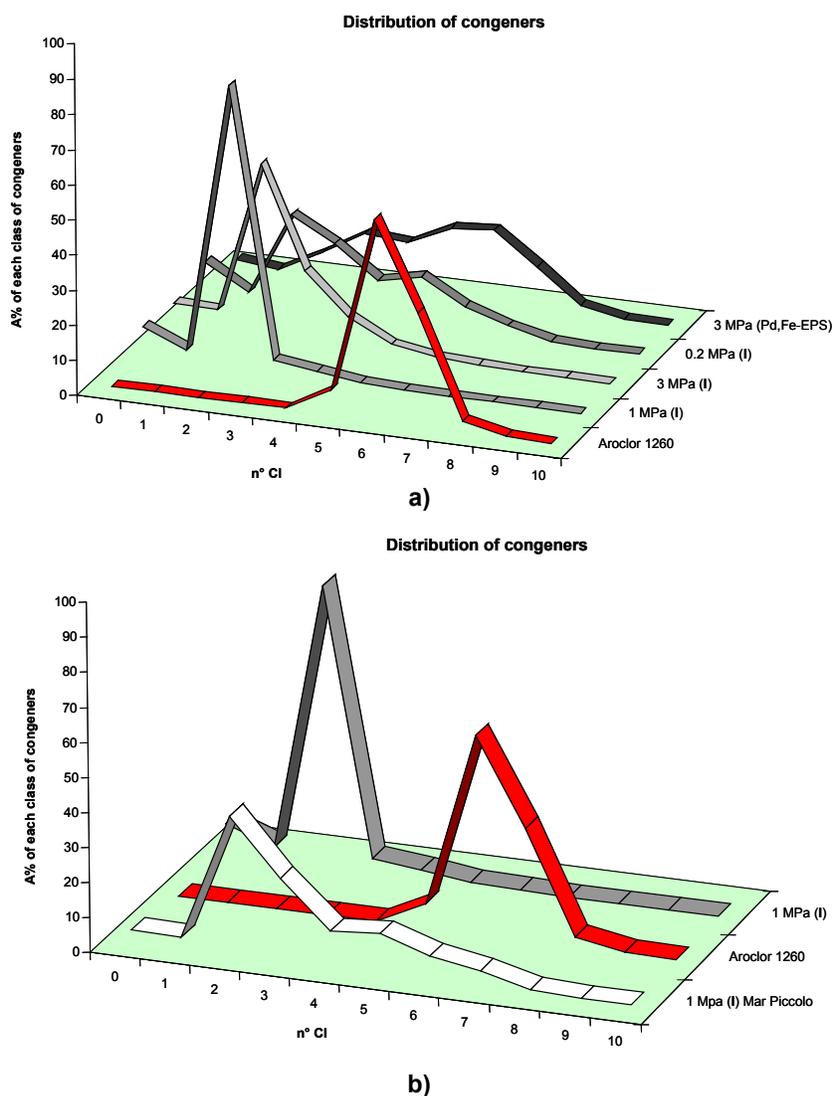


Fig. 1. Hydrodechlorination of Aroclor 1260 results compared with previous best results [5]: a) in pure water; b) diluting Aroclor 1260 in polluted sea water of Mar Piccolo

Table 2. PCBs concentrations of some dioxin-like PCBs congeners: BZ 118, 156, 157, 167, 170, 180, 189 in Aroclor 1260 and in the product obtained after hydrogenation in the presence of Pd, Fe-EPS or Met_x-EPS(I)

BZ number of some dangerous PCBs congeners ^a	Aroclor 1260 A %	Pd,Fe-EPS A % (3 MPa of H ₂) ^b	Met _x -EPS(I) A % (1 MPa of H ₂)
118 (+149)	10.18	5.16	0.04
156 (+173)	1.84	0.39	nd
157 (+197)	0.31	0.07	nd
167	0.72	0.29	nd
170 (+190)	4.25	1.33	nd
180	8.88	3.27	nd
189	0.13	0.02	nd

^a congeners number according to Ballschmiter K, Zell M(1980) Analysis of Polychlorinated Biphenyls (PCB) by Glass Capillary Gas Chromatography. Fresenius Z Anal Chem 302: 20-31; ^b see ref. [5]; nd = not detected

4. CONCLUSION

The present work shows that our idea of preparing a new polymetallic catalyst from spent catalytic converters, combining a chemical treatment and the following action of a plurimetal resistant microorganism *Klebsiella oxytoca* DSM 29614, was realized efficiently.

Positive synergies among the different metals, embedded in a peculiar polysaccharide moiety, were found when Metx-EPS (I) was used as a catalyst in the removal of PCBs by hydrodechlorination reaction, permitting to work under milder conditions with a better efficacy than our previous studies where mono- or bimetallic Pd species prepared by the same microorganism had been used [5]. This result was not easily predictable, especially when the reaction was carried out in the presence of a sample of real water taken from a contaminated site.

So this catalyst seems quite robust and potentially useful to be applied for the remediation of water taken from polluted areas.

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COMPETING INTERESTS

Authors have declared that no competing interests exist.

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APPENDIX

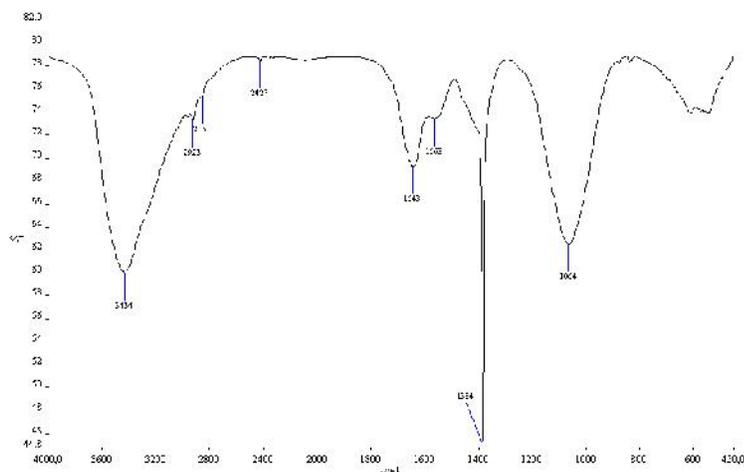


Fig. 1. FT-IR spectrum of Met_x-EPS

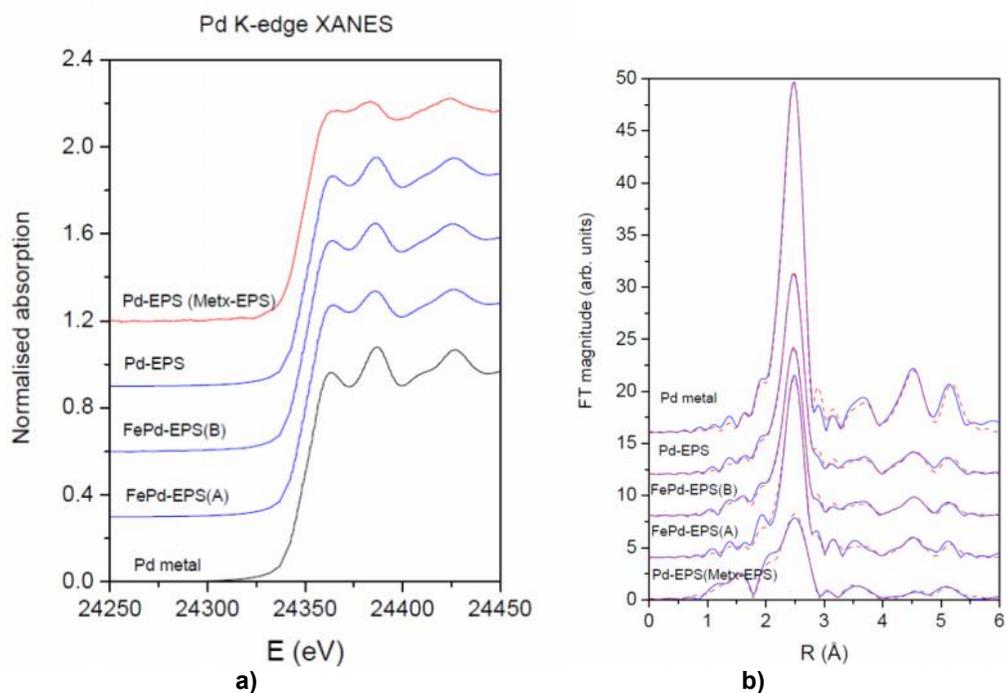


Fig. 2. a) Pd K-edge XANES spectra of the Pd-species in Met_x-EPS sample compared to Pd-EPS, FePd-EPS(A), FePd-EPS(B) samples and reference Pd metal foil with f.c.c. crystal structure; b) Pd EXAFS Fourier transform magnitude of k³-weighted Pd EXAFS spectra of Pd-EPS, FePd-EPS(A), FePd-EPS(B), Pd-species in Met_x-EPS samples and Pd metal foil, calculated in the k range 3–16 Å⁻¹ and R range 1–5.5 Å. Experiment: solid line; best-fit EXAFS model of the nearest coordination shells: dashed line

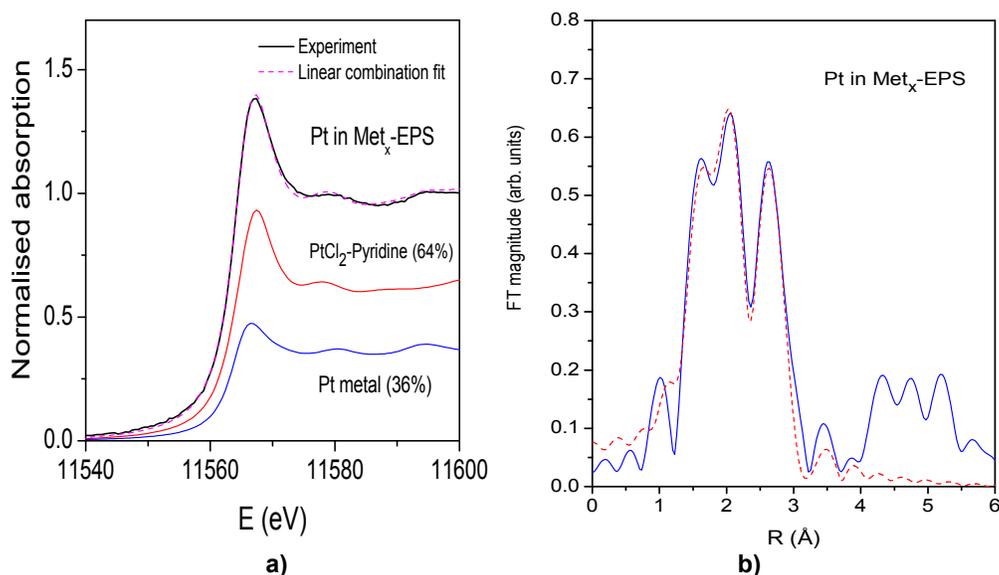


Fig. 3. a) Pt L3-edge XANES spectrum measured on the Pt in Met_x-EPS sample: Black solid line: experiment; magenta dashed line: best-fit linear combination of XANES profiles of Pt fcc metal (36%), and cis-dichlorobis pyridine platinum reference compound (64%); b) Fourier transform magnitude of k²-weighted Pt L3-edge EXAFS spectra of Pt in Met_x-EPS sample, calculated in the k range 3–12 Å⁻¹ and R range 1–3.8 Å. Experiment: solid line; best-fit EXAFS model of the nearest coordination shells: dashed line.

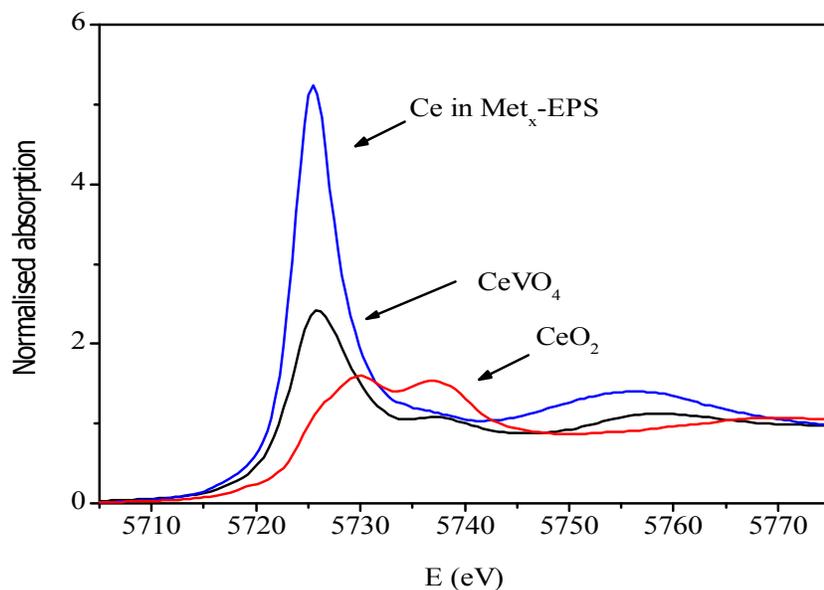


Fig. 4. Ce L3-edge XANES spectra of the Ce species in Met_x-EPS sample compared to CeVO₄ and CeO₂ spectra as references for Ce³⁺ and Ce⁴⁺, respectively.

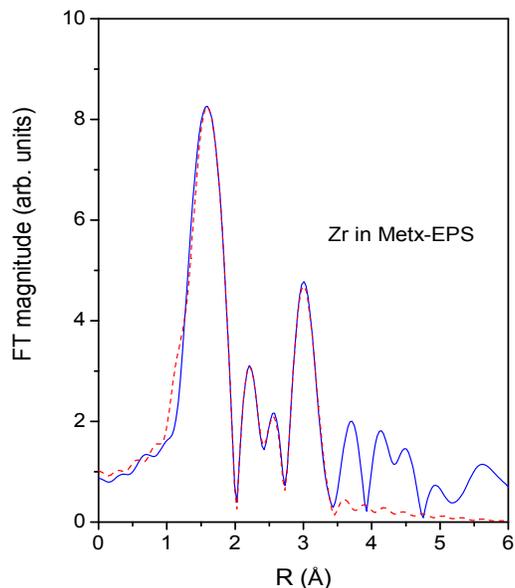


Fig. 5. Fourier transform magnitude of k₃-weighted Zr K-edge EXAFS spectra of Zr in Met_x-EPS sample, calculated in the k range 3–11 Å⁻¹ and R range 1–3.4 Å. Experiment: solid line; best-fit EXAFS model of the nearest coordination shells: dashed line.

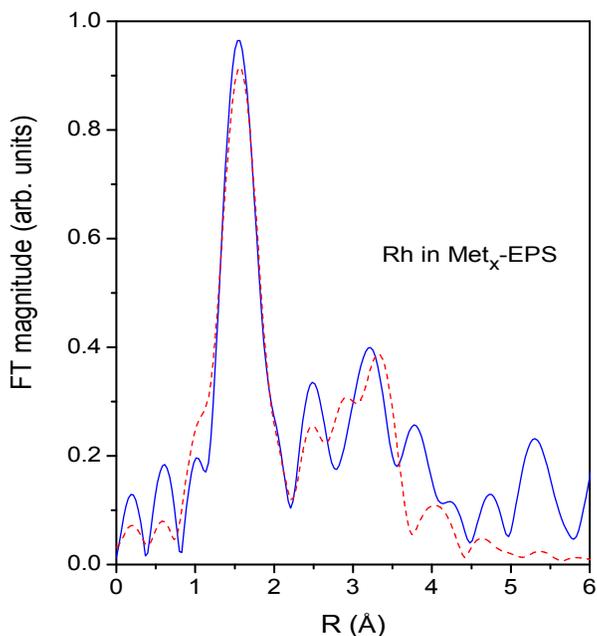


Fig. 6. Fourier transform magnitude of k₂-weighted Rh K-edge EXAFS spectra of Rh in Met_x-EPS sample, calculated in the k range 3–11 Å⁻¹ and R range 1–3.4 Å. Experiment: solid line; best-fit EXAFS model of the nearest coordination shells: dashed line.

Table 1A. Parameters of the nearest coordination shells around Pd atoms in Met_x-EPS sample and in reference Pd metal foil with fcc crystal structure (a=3.8900 Å): average number of neighbour atoms (*N*), distance (*R*), and Debye-Waller factor (σ^2). Uncertainty of the last digit is given in parentheses. A best fit is obtained with the amplitude reduction factor $S_0^2 = 0.87$. The goodness-of-fit parameter, *R*-factor, is given in the last column

Pd neigh.	<i>N</i>	<i>R</i> [Å]	σ^2 [Å ²]	<i>R</i> -factor
Pd in the Met_x-EPS sample				
O	1.4(2)	1.97(1)	0.005(2)	0.014
Pd	0.6(3)	3.50(4)	0.008(4)	
Pd	5.3(6)	2.749(4)	0.0075(6)	
Pd	4(2)	3.91(1)	0.013(1)	
Pd	7(2)	4.77(1)	0.013(1)	
Pd	7(3)	5.53(1)	0.013(1)	
Reference Pd metal foil				
Pd	12	2.745(2)	0.0058(2)	0.003
Pd	6	3.896(2)	0.0086(5)	
Pd	24	4.772(2)	0.0088(5)	
Pd	12	5.510(2)	0.0090(5)	

Table 2A. Parameters of the nearest coordination shells around Pt atoms in Met_x-EPS sample: average number of neighbour atoms (*N*), distance (*R*), and Debye-Waller factor (σ^2). Uncertainty of the last digit is given in parentheses. A best fit is obtained with the amplitude reduction factor $S_0^2 = 0.91$. The goodness-of-fit parameter, *R*-factor, is given in the last column

Pt neigh.	<i>N</i>	<i>R</i> [Å]	σ^2 [Å ²]	<i>R</i> -factor
N	1.2(3)	2.04(3)	0.002(1)	0.02
Cl	1.2(3)	2.37(4)	0.003(4)	
Pt	7(3)	2.71(1)	0.013(1)	
C	4.6(8)	2.97(7)	0.003(2)	

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